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The amphipod Corophium multisetsosum (Corophiidae) in Ria de Aveiro (NW Portugal). II. Abundance, biomass and production

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Abstract The abundance and biomass of Corophium multisetsosum Stock, 1952 were determined from benthic corer samples collected monthly over 1 yr in the upper reaches of Canal de Mira (Ria de Aveiro, Portugal). Both density and biomass over the sampling period were negatively correlated with water temperature and positively correlated with chlorophyll a concentration in the sediment. C. multisetsosum density was significantly negatively correlated with plant biomass and positively correlated with salinity. The nature of the sediment, favourable environmental conditions, high availability of food and low interspecific competition allowed the population to reach a maximal density of $200 \times 10^3$ individuals m$^{-2}$ and a maximal biomass (ash-free dry wt, AFDW) of $62 \ g_{AFDW} \ m^{-2}$. The population was highly productive, especially during the autumn/winter period. Production, estimated by two different methods (Hynes method: $251 \ g_{AFDW} \ m^{-2} \ yr^{-1}$; Morin–Bourassa method: $308 \ g_{DW} \ m^{-2} \ yr^{-1}$), was much higher than the values reported for other Corophium species. The annual P:B ratio (10) was high, but similar to values reported for Swedish populations of C. volutator and lower than the values estimated from Mediterranean populations of C. insidiosum.

Introduction

Corophium Latreille, 1806 is a cosmopolitan genus in temperate and tropical waters, and is represented by 58 species, mostly shallow marine, and often found in estuaries, harbours and in European and Ponto Caspian rivers (Barnard and Karaman 1991). Corophium species are burrow- or tube-dwellers, feeding preferentially on detritus and epipelagic microalgae, and can establish populations of extremely high densities in intertidal and subtidal areas. These amphipods are considered a key food source in mudflats used as foraging areas by migrating shorebirds, and are also an important part of the diet of fishes in sheltered coastal habitats that act as nursery and feeding grounds (Segestrelle 1959; Goss-Custard 1977; Peer et al. 1986; Wilson 1989; Matthews et al. 1992).

In Europe, the most widespread and well studied species are Corophium volutator, C. arenarium and C. insidiosum. C. multisetsosum was described by Stock (1952) from Dutch brackish waters and, since then, it has been reported from European brackish waters of the Baltic and Atlantic coast (Janta 1995). In Portugal it is known from the Ria de Aveiro, the Mondego and Sado estuaries, and also from Ria Formosa, the southern limit of its distribution (Marques and Bellan-Santini 1985, 1986, 1990; Queiroga 1990; Cunha and Moreira 1995). This species preferentially inhabits low-salinity waters (0 to 20 psu) and sandy sediments (250 to 1000 µm), where it establishes populations reaching 200 000 individuals m$^{-2}$ (Baltic Sea: Janta 1995; Ria de Aveiro: Cunha and Moreira 1995). This amphipod is often found coexisting with the gastropod Potamopyrgus jenkinsi and the amphipod Gammarus chevreuxi. C. multisetsosum is an abundant species in Ria de Aveiro, and its distribution in one of the main channels (Canal de Mira) was reported by Queiroga (1990). The species is also known from the other channels of the Ria, especially in shallow low-salinity areas. In higher-salinity areas it occurs mainly on intertidal sandy sediments near mean high water level (Cunha unpublished data).
In this study, the fluctuations in density and biomass of *Corophium multitesselatum* over 1 yr period are discussed in relation to the temporal variation of some environmental factors. Its production in a shallow area at the upper reaches of Canal de Mira is estimated and compared to values for other *Corophium* populations.

**Material and methods**

**Study area**

The area investigated (Areão) in the Ria de Aveiro is located in the upper reaches of Canal de Mira (Fig. 1). The morphodynamics, circulation patterns and environmental gradients in Canal de Mira have been described by several authors (Teles et al. 1990; Moreira et al. 1993; Corrochano et al. 1997).

In Areão, the channel is narrow and shallow (average depth <0.5 m at low water), with small intertidal areas (0 to 1 m) and extensive macrophyte meadows (mainly *Potamogeton* spp. and *Myriophyllum* spp.) that almost completely cover the channel bed.

![Fig. 1 Map of Ria de Aveiro and location of study area](image)

The maximal tidal range is about 1 m at spring tides. During the sampling period, the salinity ranged from 0 to an extreme of 15, but usually varied between 1 and 3. Water temperature ranged from 8 °C (early morning during winter) to 25 °C (afternoon during summer). Sediment type varied from medium sand (~75% of particles in the 250 to 500 μm range) in the central channel to muddy sand in the banks. Fine particles (< 63 μm) represented ~7% of the sediment dry weight, and the average organic content was 3.5%.

[For further details on sampling and laboratory procedures see study of Cunha and Moreira (1995), who also provide complete description of the macrobenthic community structure and temporal (tidal, diurnal and seasonal) variation of several environmental factors (salinity, temperature, dissolved oxygen, chlorophyll *a*, micro-seston dry weight, current velocity and plant biomass) in Areão.]

**Sampling and laboratory procedures**

Sampling was carried out monthly between May 1988 and April 1989 at low water of new-moon spring tides. Ten core samples (10 × 0.01 m²; 20 cm depth) were taken randomly in a 500 m² area bounded by the channel banks and including both subtidal and intertidal areas. The samples were sieved in the field through a 0.5 mm mesh and preserved in 10% formalin. *Corophium multitesselatum* specimens were separated from the remaining macrobenthos and kept in 70% ethanol for further examination.

All collected individuals were counted, and undamaged specimens over a wide size range were selected for measurement of head length (Lₜ, dorsal view, from tip of rostrum to posterior margin of head) and individual weight (w). Individual weights were recorded to the nearest 0.001 mg by means of a CAHN electrobalance. The total biomass (ash-free dry weight, AFDW) of each replicate in the monthly samples was determined to the nearest 0.1 mg by ignition of the dried specimens in a muffle-oven for 2 h at 450 °C.

**Data analysis**

The non-parametric Kendall’s rank correlation coefficient was used to determine the association of density and biomass of *Corophium multitesselatum* with several environmental variables recorded monthly in the study area. This statistical analysis is appropriate for variables not normally distributed. Abundance data of *C. multitesselatum* and other species present in the study area were normalised by a log(n+1)-transformation to allow a Pearson product-moment correlation analysis. Details on environmental variables and species are given by Cunha and Moreira (1995).

The relationship between the ash-free dry weight (AFDW, mg) and head length (Lₜ, mm) of *Corophium multitesselatum* was established by the equation:

\[ \text{AFDW} = a \times Lₜ^b \]  
(1)

and its physiological condition was assessed using a condition index (CI) determined by the equation:

\[ \text{CI} = \frac{w_j}{Lₜ^b} \]  
(2)

where \( w_j \) = average individual weight (mgAFDW), \( Lₜ \) = average head length (mm), and \( b \) = exponent of the allometric equation (1).

Annual production (P) was estimated using the Hynes size-frequency method modified by Menzies (1980):

\[ P = \left[ i \sum (d_j - d_{j+1}) \times \sqrt{(w_j \times w_{j+1})} \right] \frac{12}{CPI} \]  
(3)

where \( i \) = number of size classes; \( d_j \) = mean density in size class \( j \) (individuals m⁻²); \( w_j \) = mean individual weight in size class \( j \) (mgAFDW); and \( CPI \) = cohort production interval (mo); and the multiple regression equation of Morin and Bourassa (1992):

\[ \log P = -0.75 + 1.01 \log B - 0.34 \log M + 0.037T \]  
(4)

where \( B \) = mean annual biomass (gAFDW m⁻²); \( M \) = mean individual weight (gAFDW) = \( BD^{-1} \); \( D \) = mean annual density; \( DW \) = dry biomass.
weight; \( T^* = \frac{\sum B_i T_i}{\sum B_i} \); \( B_i \) = biomass in sample \( i \); and \( T_i \) = water temperature in sample \( i \).

With the first method, the mean individual ash-free dry weight in each size class \( \bar{w}_i \) is obtained from an allometric equation, while with the second method the mean individual dry weight \( \bar{M} \) is obtained from the actually measured biomass for each monthly sample \( (B_i) \).

**Results**

Temporal variation of density and biomass

*Corophium multisetosum* maintained extremely high densities all year round in the study area (Fig. 2). Water temperature and salinity fluctuations, depending mainly on the climatic conditions (e.g. insolation and rainfall), were the chief abiotic factors related to population fluctuations of *Corophium multisetosum* in Areão. A small peak of density and biomass occurred at the end of spring \( (D = 52.3 \times 10^3 \text{ individuals m}^{-2} \) in June, \( B = 16.7 \text{ g m}^{-2} \) in July). A more effective period of increasing abundance and biomass occurred during the autumn, when temperature decreased from 23 to 18 °C, and maxima were reached in winter \( (D = 199.6 \times 10^3 \text{ individuals m}^{-2} \) in December, \( B = 62.0 \text{ g m}^{-2} \) in February). A gradual decrease in abundance and biomass occurred at low water temperatures \( (10 \text{ to } 12 \degree \text{C}) \). Summer minima \( (D = 21.6 \times 10^3 \text{ individuals m}^{-2} \) in August, \( B = 7.1 \text{ g m}^{-2} \) in September) occurred at the highest temperatures \( (=24 \degree \text{C}) \). Increasing trends in abundance and biomass were observed when salinity varied between 1 and 3 psu. Minima \( (D = 22.4 \times 10^3 \text{ individuals m}^{-2}, B = 5.4 \text{ g m}^{-2} \) in May) occurred at very low salinities \( (< 0.5 \text{ psu}) \), during rainy periods.

Figure 2 illustrates the fluctuations of the *Corophium multisetosum* population (abundance and biomass) and water temperature and chlorophyll concentration in the sediment, over the study period. The trends observed in chlorophyll concentration, density and biomass are similar, but occurred successively, with a 1 to 2 mo lag. During summer and autumn, an increasing concentration of chlorophyll was accompanied by increasing density and biomass of *Corophium multisetosum*. During winter, chlorophyll concentrations maintained maximal values \( (=6 \text{ mg m}^{-2}) \) and, although the temperature was no longer favourable to reproduction (reflected by decreasing density), biomass continued to increase until February through individual growth.

Figure 3 shows the physiological condition \( (\text{Cl} = \bar{w}_i/L_i^{3.241}) \) of *Corophium multisetosum* over the study period (Fig. 3A) and its relation to chlorophyll concentration in the sediment (Fig. 3B). An increase in chlorophyll concentration from June to November was accompanied by an improvement in the physiological condition of the amphipod. Condition index \( (\text{Cl}) \) was 1.5 in June; this increased gradually until it attained twice this value in December.

Both density and biomass were significantly correlated with chlorophyll in the sediment (positively) and with water temperature (negatively) (Table 1). Moreover, biomass was highly correlated with the concentration of chlorophyll observed 1, 2 and 3 mo earlier \( (r = 0.639, 0.689, 0.611, \text{ respectively}; P < 0.005) \). This variable was also highly correlated with temperature \( (r = -0.469 \text{ and } -0.455 \text{ for the correlation of chlorophyll with temperature at high and low water, respectively; } P < 0.025) \).

The temporal variation in density of *Corophium multisetosum* was positively correlated with salinity at low water and negatively correlated with plant biomass (mainly *Potamogeton* spp. and *Myriophyllum* spp.) and with oxygen saturation at low water (Table 1). The latter two factors were closely related to each other \( (r = 0.489; P < 0.025) \). Plant biomass displayed peak values in August and March and minimum values from November to February.

There was a significant positive correlation between biomass of *Corophium multisetosum* and both chlorophyll concentration and dissolved oxygen recorded at high water. The highest concentrations of chlorophyll in the water column were recorded from January to March \( (190 \text{ to } 290 \text{ mg m}^{-3}) \), but a smaller peak occurred in September \( (130 \text{ mg m}^{-3}) \).

**Interaction with other species**

The abundance of *Corophium multisetosum* throughout the study period was significantly correlated with the abundance of only a few species. There was a positive correlation with the abundance of other crustaceans in the area (Gammarus chevreuxi, Leptocheirus pilosus, Cyathura carinata and Lekanesphaera hookeri) all of which showed quite similar trends in seasonal variation. Significant negative correlations were recorded between
Fig. 3 A Corophium multitetosum. Changes in individual weight and head length over study period (May 1988 to April 1989); B relationship between physiological condition of *C. multitetosum* and chlorophyll *a* concentration in sediment.

*Corophium multitetosum* and the polychaete *Nereis diversicolor*, the oligochaete *Ophidion serpentina*, and the gastropods *Valvata piscinalis* and *Physa acuta*, all of which occurred at relatively low densities (50 to 260 individuals m$^{-2}$). However, the correlation of *C. multitetosum* abundance with that of the most abundant species, such as the gastropod *Potamopyrgus jenkinsi* (103 × 10$^3$ individuals m$^{-2}$), and the oligochaete *Limnodrilus hoffmeisteri* (8 × 10$^3$ individuals m$^{-2}$), was not significant.

Table 1 Kendall’s rank correlation ($r$) between density (individuals m$^{-2}$) and biomass, as ash-free dry wt ($g_{AFDW}$ m$^{-2}$) of *Corophium multitetosum* and several environmental factors recorded throughout 1 yr period (*HW* high water, *LW* low water; *DW* dry wt; *P* < 0.05; **P** < 0.025; ***P*** < 0.005 ($n = 12$, one-tailed test))

<table>
<thead>
<tr>
<th>Variable</th>
<th>Variable range</th>
<th>Kendall’s correlation</th>
</tr>
</thead>
<tbody>
<tr>
<td>Vegetation biomass</td>
<td>(146–289 g$_{AFDW}$ m$^{-2}$)</td>
<td>−0.504 **  −0.351 NS</td>
</tr>
<tr>
<td>Sediment Chlorophyll $a$</td>
<td>(1.8–5.9 mg m$^{-2}$)</td>
<td>0.576 ***  0.485 **</td>
</tr>
<tr>
<td>Organic content</td>
<td>(1.3–10.2%)</td>
<td>−0.030 NS  0.000 NS</td>
</tr>
<tr>
<td>Water Salinity (HW)</td>
<td>(0.5–6.5)</td>
<td>0.362 NS  0.268 NS</td>
</tr>
<tr>
<td>(LW)</td>
<td>(0.2–2.2)</td>
<td>0.485 **  0.291 NS</td>
</tr>
<tr>
<td>Temperature (HW)</td>
<td>(12.4–25.3 °C)</td>
<td>−0.406 *  −0.594 ***</td>
</tr>
<tr>
<td>(LW)</td>
<td>(11.0–23.5 °C)</td>
<td>−0.394 *  −0.606 ***</td>
</tr>
<tr>
<td>Dissolved O$_2$ (HW)</td>
<td>(9.0–18.0 mg l$^{-1}$)</td>
<td>0.137 NS  0.473 **</td>
</tr>
<tr>
<td>(LW)</td>
<td>(5.0–13.8 mg l$^{-1}$)</td>
<td>−0.061 NS  0.229 NS</td>
</tr>
<tr>
<td>O$_2$ saturation (HW)</td>
<td>(95–188%)</td>
<td>−0.091 NS  0.182 NS</td>
</tr>
<tr>
<td>(LW)</td>
<td>(57–138%)</td>
<td>−0.455 **  −0.152 NS</td>
</tr>
<tr>
<td>Chlorophyll $a$ (HW)</td>
<td>(6–288 mg m$^{-3}$)</td>
<td>0.273 NS  0.485 **</td>
</tr>
<tr>
<td>(LW)</td>
<td>(3–30 mg m$^{-3}$)</td>
<td>−0.091 NS  0.121 NS</td>
</tr>
<tr>
<td>Micro-seston (HW)</td>
<td>(2–56 g$_{DW}$ m$^{-3}$)</td>
<td>0.055 NS  0.273 NS</td>
</tr>
<tr>
<td>(LW)</td>
<td>(1–33 g$_{DW}$ m$^{-3}$)</td>
<td>0.200 NS  0.200 NS</td>
</tr>
</tbody>
</table>

Production

The Hynes and Morin–Bourassa methods were used to estimate the secondary production of *Corophium multitetosum* since it was not possible to accurately discriminate the different overlapping cohorts (Cunha et al. 2000b).

For the determination of the mean annual cohort (Hynes method) individuals were divided into 12 size-classes (see also Cunha et al. 2000b, for demographic structure of the population). The mean individual weight of each size-class was obtained from the equation expressing the relationship (Fig. 4) between the ash-free dry weight (AFDW, mg) and head length ($L_h$, in mm):
AFDW = 1.7671 × Lₜ₉³.²₄₁
(n = 234; r = 0.965, P > 0.001;
0.183 ≤ Lₙ ≤ 1.000) ,

where the exponent was significantly higher than 3
(lₚₙ = 4.47; 232 df; P < 0.001). The ash-free dry weight
can be converted to dry weight using the equation:

DW = 1.129 × AFDW ,

where in the value 1.129 was obtained from the data on
mean percentage of ashes for Corophium multisetosum.

A correction factor (CPI) of 6 mo, corresponding to
the mean longevity of Corophium multisetosum in Areão
(Cunha et al. 2000b) was used. The estimate of pro-
duction using the Hynes method, 250.95 gAFDW
m⁻² yr⁻¹ (Table 2), was lower than that obtained by the
Morin–Bourassa method, 308.35 gDW m⁻² yr⁻¹
(=273.02 gAFDW m⁻² yr⁻¹). The annual P:B ratio of
Corophium multisetosum in the study area was high, and similar
estimates (Table 2) were obtained by the two methods
(10.14 and 9.92 for Hynes and Morin–Bourassa meth-
ods, respectively).

Seasonal estimates showed that the differences be-
tween the two methods were more important for the
autumn/winter season. During this period, the mean
density, mean biomass, production and P:B ratio were
higher than in the spring/summer period (Table 2). The
autumn/winter population contributed ≈70% of the
annual production of Corophium multisetosum in Areão.

<table>
<thead>
<tr>
<th>Method, Period</th>
<th>D</th>
<th>B</th>
<th>M</th>
<th>Over 6 mo:</th>
<th>Over 1 yr:</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>P</td>
<td>P:B</td>
<td>P</td>
<td>P:B</td>
<td></td>
</tr>
<tr>
<td>Hynes (12 size-classes, AFDW)</td>
<td>56.25</td>
<td>4.70</td>
<td>250.95</td>
<td>10.14</td>
<td></td>
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<tr>
<td>Sp–Su</td>
<td>38788</td>
<td>16.21</td>
<td>76.25</td>
<td>4.70</td>
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</tr>
<tr>
<td>Au–Wi</td>
<td>130022</td>
<td>33.30</td>
<td>174.69</td>
<td>5.25</td>
<td></td>
</tr>
<tr>
<td>Annual</td>
<td>84405</td>
<td>24.76</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Morin–Bourassa (Tᵣ = 15.16, DW)</td>
<td>83.16</td>
<td>4.48</td>
<td>308.35</td>
<td>9.92</td>
<td></td>
</tr>
<tr>
<td>Sp–Su</td>
<td>38788</td>
<td>18.55</td>
<td>415 × 10⁻⁶</td>
<td>83.16</td>
<td>4.48</td>
</tr>
<tr>
<td>Au–Wi</td>
<td>130022</td>
<td>43.60</td>
<td>335 × 10⁻⁶</td>
<td>222.45</td>
<td>5.10</td>
</tr>
<tr>
<td>Annual</td>
<td>84405</td>
<td>31.07</td>
<td>368 × 10⁻⁶</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>
**Discussion**

Environmental factors

Previous experiments on salinity and substrate selection (Queiróga 1990; Chatwin 1991) showed that *Corophium multisetosum* prefers salinities from 2.5 to 10 psu and avoids the finest (63 μm) and coarsest (2 mm) grades over a range of substrate particle-sizes. These results correspond to the observed field data for Areeão, and may explain the remarkable density of *C. multisetosum* in the area. Further observations on the local distribution of *C. multisetosum* in neighbouring areas (±1 km) also showed that density decreased both upstream and seawards. *C. multisetosum* appeared to avoid low-flow areas and substrates with a high percentage of fine particles, and it was usually absent from sediments with a black reduced layer. Laboratory experiments (Ré 1996) showed that in sediments consisting of >50% of fine particles (<0.065 mm), reproduction of *C. multisetosum* was inhibited and construction of tubes became difficult. In the presence of higher percentages of fine particles, the amphipod just burrows in the sediment.

The available literature on factors affecting the distribution and abundance of *Corophium* species is vast, although most studies focus on *C. volutator* and *C. arenarium*. These species are highly tolerant to salinity variations, and thus the nature of the substratum often determines their distribution (Gee 1961; Meadows 1964a, b, c). However, McLusky (1968) suggested that the abundance and distribution of *C. volutator* in the Ythan estuary, Scotland, were controlled by a critical minimum salinity. The positive correlation observed between *C. multisetosum* abundance and salinity at low water in Areeão is in accordance with the results of salinity-preference experiments which showed that very low salinities (>0.0 psu) were rejected (Queiróga 1990). Further laboratory experiments (Ré 1996) showed that at very low salinities (<0.3 psu; T = 22 °C) survivorship is significantly reduced, while intermediate values (> 2 to <18 psu) favour survivorship, growth and reproduction. Previous field data on *C. multisetosum* (Queiróga 1990) demonstrated that extremely low salinities were associated with severe decreases in abundance. In March 1986, when salinity remained close to 0.0 psu at both low and high water, the estimated density of *C. multisetosum* in Areeão was 827 individuals m⁻², >100 times less than in December 85 (D' in Fig. 2). Since tides have little influence in Areeão, rainfall is the main factor affecting salinity. Drastic decreases in *C. multisetosum* abundance may be associated with increased freshwater inflow following rainy periods on the stream catchment area. Year-to-year density fluctuations may occur, depending on rainfall. Casabianca (1972/1973) reported a mortality rate of ~95% in *C. insidiosum* from Biguglia, Corsica, caused by decreased salinity and increased current velocity following exceptional rainfall.

The negative correlation between temperature and both abundance and biomass of *Corophium multisetosum* may seem unexpected. However, it must be considered that this is a cold-temperate species (Lincoln 1979), with its southern limit in Portugal. In principle, the minimum water temperature occurring in Areeão (8 °C) should not be limiting to a species that must endure much lower values in northern areas. However, long periods of high temperatures (>24 °C) may deter reproduction (Cunha et al. 2000a, b) and contribute to the summer minima of abundance and biomass. The possible physiological sensitivity of *C. multisetosum* to low oxygen concentrations may also explain the summer decline, which could be caused by either mortality or migration. Cunha and Moreira (1995) showed that, in summer, when dense vegetation and high temperatures occur, the oxygen concentrations, which reached oversaturation during the day in Areeão, are followed by a drastic drop during the night (~2 mg l⁻¹; 15 to 30% saturation). Physiological sensitivity of *Corophium* species to some factors such waste discharges, sulphide, and hypoxia has been mentioned or demonstrated by several authors (Segesträle 1959; Meadows et al. 1981; Esselink et al. 1989; Gamenick et al. 1996). However, other studies failed to establish any correlation between local distribution and abiotic factors (Murdock et al. 1986; Ölafsson and Persson 1986; Jensen and Kristensen 1990; Hughes and Gerdl 1997).

The tube-building and feeding activities of *Corophium* spp. may interact with environmental factors and processes, for instance, enhancing substrate instability, reducing the density of micro-organisms and organic content (Meadows and Tait 1989; Gerdl and Hughes 1993, 1994a, b), or stimulating sediment-oxygen respiration and denitrification (Rysgaard et al. 1995). Such interactions prevent the interpretation of causal relations to the abundance of *Corophium* spp.. This could be the case of the observed negative correlation between the temporal variation of plant biomass and *C. multisetosum* abundance in Areeão. In fact, high densities of *C. multisetosum* certainly contribute to sediment instability, disfavouring colonisation by rooted plants. However, plant roots may prevent burrowing. Murdoch et al. (1986) determined that plant detritus supplied <5% of the nutritional demands of *Corophium* spp.. However, decay of vegetation during late summer and autumn is followed by an increase in bacterial decomposition activity and, consequently, an increase in food supply for detritus-feeders. This process probably favours the growing population of *C. multisetosum* during the autumn. Substrate instability, food supply and the above mentioned relationship to oxygen concentration illustrate the complexity of the interaction between *C. multisetosum* and rooted vegetation.

*Corophium* spp. are predominantly unselective deposit-feeders, although suspension-feeding and epipsammic browsing may also occur (Fenchel et al. 1975; Nielsen and Kofod 1982). Several authors have shown that benthic microalgae, mainly diatoms, form a signif-
ificant part of their diet (Morrissey 1988a, b; Gerdol and Hughes 1994a, b; Smith et al. 1996). The importance of benthic microalgae to the nutritional requirements of *C. multisetosum* is suggested by the strong correlation between the density, biomass and physiological condition of the amphipod and the chlorophyll *a* concentration in the sediments.

The negative correlation between chlorophyll *a* concentration and temperature naturally contributes to the equally negative relationship between density and biomass of *Corophium multisetosum* and temperature.

**Interspecific relations**

In addition to the effect of environmental factors, interspecific relations, including competition, predation and amensalism, may also influence distribution and abundance of *Corophium* spp. Beukema and Flach (1995) suggested that the clear zonation of *Corophium* species in the Wadden Sea shores is determined by their physiological ability to endure prolonged emersion at the upper intertidal and by the outcome of their interaction with other species at the lower intertidal. Intensive sediment-reworking by *Arenicola marina* or high densities of *Cerastoderma edule* can be effective in reducing numbers of *Corophium* species, dissuading the amphipod from burrowing (Flach 1992, 1993; Flach and de Bruin 1993). However, data on possible negative interactions with other species such as *Macoma balthica*, *Nereis diversicolor* or *Hydrobia ulvae* are inconclusive or contradictory (Jensen 1988; Morrissey 1988a, b; Flach 1992; Jensen and Andre 1993; Hughes and Gerdol 1997).

The results from Areão showed that the remarkably high density of *Limnodrilus hoffmeisteri* and *Potamogeton jenkinsii* had no significant effect on the seasonal variation of *C. multisetosum*. Moreover, the negative correlation with some low-density freshwater species cannot be ascribed to significant competitive interference, and is probably due to opposing effects of salinity fluctuations. Positive associations with other crustacean species are probably induced by coincident fluctuations resulting from similar life-history strategies.

It has been demonstrated that size-selective predation may influence the life-history patterns and population fluctuations of *Corophium* species (Wilson 1989; Matthews et al. 1992). These amphipods may comprise a significant food resource for waders (Goss-Custard 1977; Peer et al. 1986; Jensen and Kristensen 1990), fishes, and decapods (Segestrela 1959; Wilson 1989). Although *Corophium* spp. are a substantial prey for fishes, Schneider and Harrington (1981) concluded that the number of fish predators would have to be one order of magnitude greater than the number of birds (sandpipers) to have the same impact on the prey population.

Ria de Aveiro is an important staging area for birds; waders begin to increase their density in October, reach a maximum in December, and remain until February or March. Despite the high density of invertebrates, Areão is not usually frequented by waders, except for the dunlin *Calidris alpina*, which visits the area occasionally during late spring. The absence of large intertidal flats, the existence of trees along the margins, and the distance (~20 km) to the usual roosting sites probably render the area unattractive to these birds (Luis 1998, and personal communication).

The fish community in Areão includes mainly estuarine residents (*Atherina boyeri*, *Platichthys flesus*, *Gobiidae*, *Synagathidae*), marine seasonal users (*Dicentrarchus labrax*, *Chelon labrosus*, *Liza spp.*, *Mugil cephalus*) and the catadromous migrant *Anguilla anguilla*. All these species feed on invertebrates, either strictly or in combination with other types of food (plankton, plants, detritus, other fishes). *A. anguilla* is probably the main predator of *Corophium multisetosum* in the study area because it represents 10 to 12% of the abundance and 31 to 50% of the total biomass of the fish community (Rebelo 1992).

**Production**

Production and P:B estimates of *Corophium* species from European populations are summarised in Table 3 (different methods were used in the different areas). The highest production of *C. insidiosum* (61 gDW m⁻² yr⁻¹) and *C. volutator* (95.6 gDW m⁻² yr⁻¹), observed in the Biguglia lagoon, Corsica (Casabianca 1975) and in the Thames estuary, England (see Möller and Rosenberg 1982), respectively, was much lower than the production of *C. multisetosum* in Areão (308 gDW m⁻² yr⁻¹). The lowest production (0.012 gAFDW m⁻² yr⁻¹) was estimated for *C. sextoni* in Torbay, England (Hughes 1978). The latter is a fouling species, and its production values are dependent upon the availability of suitable hard substrata.

The P:B estimates for *Corophium* species show a wide variation but, in general, are higher than for other amphipod species (Birklund 1977; Collie 1985). Variation of the P:B ratio may occur in relation to water temperature and size (Schwinghamer et al. 1986; Morin and Bourassa 1992). Gratto et al. (1983) ascribed the difference in the production and P:B ratios of two populations of *C. volutator* from the Bay of Fundy to distinct life-history patterns induced by different temperatures. The maximal P:B ratio of 19.5, recorded for *C. insidiosum* in the Biguglia lagoon (Casabianca 1975), may be explained both by the high water temperatures (30 to 34 °C) and the small size of the species (length up to ~5 mm) relative to *C. volutator* (10 mm) and *C. multisetosum* (10 mm). The P:B ratio of *C. multisetosum* (9.9 to 10.1) compared well with the highest values estimated for *C. volutator* in the Gullmarsvik (11.3) and in the Sandvik (10.4), in Sweden, using a cohort separation method (Möller and Rosenberg 1982). During the winter, these areas were ice-covered for several months. The major contribution to the production, up to 85%, was made in the summer period (June/August) when the temperature rose to ~20 °C. In
Table 3 Corophium spp. Production estimates (g m⁻² yr⁻¹). *Values from Denmark estimated over 4 mo period (May to September) DW dry wt; AFDW ash-free dry wt

<table>
<thead>
<tr>
<th>Species, Location</th>
<th>Production</th>
<th>P:B</th>
<th>Source</th>
</tr>
</thead>
<tbody>
<tr>
<td>C. insidiosum</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Denmark</td>
<td>0.23–8.37* (DW)</td>
<td>1.8–4.9</td>
<td>Birklund (1977)</td>
</tr>
<tr>
<td>Corsica</td>
<td>3–61 (DW)</td>
<td>12.0–19.5</td>
<td>Casabianca (1975)</td>
</tr>
<tr>
<td>C. curvispinum</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>The Netherlands</td>
<td>5.85–11.26 (AFDW)</td>
<td>2.7–3.5</td>
<td>Rajagopal et al. (1999)</td>
</tr>
<tr>
<td>C. sextoni</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>England</td>
<td>0.012–0.016 (AFDW)</td>
<td>2.1–2.8</td>
<td>Hughes (1978)</td>
</tr>
<tr>
<td>C. volutator</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Denmark</td>
<td>2.46–3.80* (DW)</td>
<td>3.2–4.4</td>
<td>Birklund (1977)</td>
</tr>
<tr>
<td>England</td>
<td>95.6 (DW)</td>
<td>7.7</td>
<td>Mossman (1977)*</td>
</tr>
<tr>
<td>Canada</td>
<td>8.8–22.1 (DW)</td>
<td>2.9–6.8</td>
<td>Gratto et al. (1983)</td>
</tr>
<tr>
<td>C. multisetosum</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Portugal</td>
<td>250.95 (AFDW)</td>
<td>10.1</td>
<td>This study</td>
</tr>
<tr>
<td>Portugal</td>
<td>308.35 (DW)</td>
<td>9.9</td>
<td>This study</td>
</tr>
</tbody>
</table>


Areão, the most productive period of C. multisetosum occurred during the autumn at temperatures around 15 to 18 °C. These data suggest that the weighted temperature (T') is probably a better indicator than average annual temperature when comparing P:B estimates from different populations because it accounts for the specific pattern of biomass fluctuations throughout the year.

Morin and Bourassa (1992) found that production is directly proportional to the mean annual biomass, while the P:B ratio is independent of biomass (see also Banse and Mosher 1980; Brey 1990). Our data confirm this: despite the differences in the estimates of B and P obtained by the two methods, the values of the P:B ratio coincided. The difference between the two production estimates (8%) may be explained by differences in biomass assessment: the Hynes method determines biomass by an allometric equation, while the Morin–Bourassa method uses the actual measured biomass. The mean biomass estimated by the Hynes method was 24.76, 16.21 and 33.30 gAFDW m⁻² for the 1 yr period, the spring/summer, and the autumn/winter seasons, respectively. The corresponding values obtained from direct determination of biomass were 27.51, 16.08 and 38.95, showing that the allometric regression underestimated the actual measured mean annual biomass by ∼10% and the autumn/winter mean biomass by ∼14.5%. These results indicate that the use of a single annual allometric relation is not appropriate to accurately estimating seasonal production. Because of the difference in the physiological conditions of the population during the two seasons (Fig. 3A), separate allometric relations should have been determined.

Autumn/winter was the most productive period, mainly as a result of the extremely high density and biomass of Corophium multisetosum at this time. The difference in seasonal P:B ratios reflects different demographic structures of the population during the two periods. Recruitment was higher in autumn than in spring (Cunha et al. 2000b), and the higher percentage of juveniles resulted in a smaller average size and contributed to the higher P:B ratio. The exponents that characterise the allometry of P:B with average size are negative (Schwinghamer et al. 1986; Morin and Bourassa 1992). The effect of the demographic structure in production and P:B estimates is accounted for by the use of a mean cohort in the Hynes method and by the mean individual weight (M) in the Morin–Bourassa method.

Final remarks

The nature of the sediment and the favourable environmental conditions, especially during the autumn, as well as the high availability of food resources, enable Corophium multisetosum to establish a dense and highly productive population in Areão. This population is one of the most abundant ever reported for Corophium species. A similar value of 200 × 10³ individuals m⁻² was estimated for a C. multisetosum population in the Dead Vistula, Poland (Janta 1995), while the maximal density reported for C. volutator was 120 × 10³ individuals m⁻² in Bassholm, Sweden (Möller and Rosenberg 1982).

The sediment surface occupied by each individual of Corophium volutator orientalis was determined by Omori and Tanaka (1998), who estimated the maximum densities as 602, 163, 83 and 50 × 10³ individuals m⁻² for juveniles, young, females and males, respectively. Considering that C. multisetosum and C. volutator are about the same size, the densities in Areão may be close or even above the theoretical maximal density, and the population is probably under high intraspecific competition pressure. Previous evidence from other Corophium species indicates that dense populations may experience intraspecific competition (Wilson 1989; Jensen and Kristensen 1990; Matthews et al. 1992). In such cases, swimming activity usually increases and high mortality rates or emigration of juveniles may occur (Jensen and Kristensen 1990). The implications of the swimming
behaviour of *C. volutator* in dispersal and colonisation have been discussed by Hughes (1988) (see also Essink et al. 1989; Hughes and Horsfall 1990).

In a population open to migration, production is lost to, and gained from, other areas. Because of the extremely high densities of *Corophium multisetosum* in Areão, especially during the autumn, intra-specific competition inducing swimming activity and migration may occur, and an important part of the production is probably exported to other areas. Further research should address the hypothesis that *C. multisetosum* in Areão constitutes a stock population for the colonisation of other areas in Ria de Aveiro. Quantitative data on the pelagic occurrence and dispersal of *C. multisetosum* are needed to assess the importance of this population in energetic and genetic fluxes within the Ria.


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